Oak Barrens

Overview: Oak barrens is a fire-dependent, savanna type dominated by oaks, having between 5 and 60 percent canopy, with or without a shrub layer. The predominantly graminoid ground layer is composed of species associated with both prairie and forest communities. Oak barrens are found on droughty soils and occur typically on nearly level to slightly undulating sandy glacial outwash, and less often on sandy moraines or ice contact features.

Global and State Rank: G2?/S1

Range: Barrens and prairie communities reached their maximum coverage in Michigan approximately 4,000-6,000 years before present, when postglacial climatic conditions were comparatively warm and dry. During this time, xerothermic conditions allowed for the invasion of fire-dependent, xeric vegetation types into a large portion of the Lower Peninsula and into sections of the Upper Peninsula. With the subsequent shift of more mesic climatic conditions southward, there has been a recolonization of mesic vegetation throughout Michigan. The distribution of fire-dominated communities, such as oak barrens, has been reduced typically to isolated patches concentrated along the climatic tension zone. In the 1800s, oak barrens were located in the interior of the southeastern Lower Peninsula on sandy glacial outwash and coarse-textured moraines (Comer et al. 1995). Presently the distribution of this community has been reduced to degraded remnants throughout its original range. In addition to southern Michigan, oak barrens remnants occur south of the tension zone through Wisconsin and southeastern Minnesota, and in the glaciated portions of Ohio, Indiana, Illinois, Iowa, Missouri, Kansas, and Nebraska (Chapman et al. 1995, NatureServe 2001).

Rank Justification: At the time of European settlement, oak savanna communities covered some 11-13 million hectares of the Midwest. Presently oak savanna remnants occur on just 0.02% of their presettlement extent (Nuzzo 1986). The notes of the original land surveyors of
Ecoregional map of Michigan (Albert 1995) depicting historical distribution of oak barrens (Albert et al. 2008)
Michigan reveal that in the 1800s, oak barrens covered approximately 719,042 acres or 1.9% of the state, distributed patchily across the four lower tiers of counties. In Allegan and Ottawa Counties, savanna communities were typically oak-pine barrens, while lakeplain oak openings were prevalent in the thumb region (Huron, Bay, and Tuscola Counties). Oak openings shared the same range as oak barrens but occurred on dry-mesic to mesic soils as opposed to droughty sites. Surveyors’ notes indicate that high concentrations of oak barrens occurred in the following counties: Oakland County (28% or 200,557 acres), Jackson County (12% or 84,204 acres), Livingston County (11% or 81,176), and Washtenaw County (9% or 62,966 acres). Today merely a few hundred acres of oak barrens remain in Michigan with small, restorable remnants occurring in Cass, Branch, Livingston, Jackson, Washtenaw, and Van Buren Counties. This rare community constitutes less than 0.0005% of the present vegetation of Michigan.

Oak barrens have been cleared for sand mining, agriculture, and residential and urban development (Chapman et al. 1995). Alteration of historic fire regimes has shifted most barrens types into woodlands and forest (Curtis 1959, Faber-Langendoen 1993). Wildfire suppression policies instituted in the 1920s in concert with road construction, expansion of towns, and increased agriculture caused a dramatic decrease in fire frequency and intensity (Abrams 1992). The reduction of fire in the landscape resulted in the succession of open oak barrens to closed-canopy forests dominated by black and white oaks with little advanced regeneration of oaks and a vanishing graminoid component (Chapman et al. 1995). In addition, timber exploitation of oaks in the 1920s destroyed or degraded oak barrens across Michigan (Michigan Natural Features Inventory 1995). Many oak barrens fragments are currently completely dominated by black oak as the result of selective harvest of canopy white oak (Minc and Albert 1990). In addition to simplified overstory structure, these communities are often depauperate in floristic diversity as the result of fire suppression and subsequent woody encroachment, livestock grazing, off-road vehicle activity, and the invasion of exotic species (Michigan Natural Features Inventory 1995).

**Landscape and Abiotic Context:** Oak barrens occur on well-drained, nearly level to slightly undulating sandy, glacial outwash, and less often on sandy moraines or ice contact features. Oak barrens typically occur in the driest landscape positions, such as ridge tops, steep slopes, south and west facing slopes, and flat sand plains. This xeric, fire-prone community is characterized by soils that are infertile, coarse-textured, well-drained sand or loamy sand with medium to slightly acid pH and low water retaining capacity. Soils contain low organic matter and lack the fine-textured illuvial horizon associated with soils of the oak openings and are thus driest. Oak barrens and oak-pine barrens typically occur in bands surrounding prairie (Chapman et al. 1995, Kost et al. 2007).

Oak barrens are distributed in Michigan’s Region I, Southern Lower Michigan (Albert et al. 1986). This region has a warm, temperate, rainy to cool, snow-forest climate with hot summers and no dry season. The daily maximum temperature in July ranges from 29°C to 32°C (85°F to 90°F) and the daily minimum temperature in January ranges from -9°C to -4°C (15°F to 25°F). The number of freeze-free days is between 120 and 220, and the average number of days per year with snow cover of 2.5 cm or more is between 10 and 60. The mean annual total precipitation for Region I is 820 mm (Albert et al. 1986, Barnes 1991).

**Natural Processes:** Curtis (1959) suggested that oak barrens originated when prairie fires spread into surrounding closed oak forest with enough intensity to create open barrens. Repeated low intensity fires working in concert with drought then maintain these barrens (Curtis 1959, Faber-Langendoen and Tester 1993). Oak barrens persist when fire disturbance and/or drought prevents canopy closure and the dominance of woody vegetation. Presently, the prevalent catalyst of fires is lighting strike but historically, Native Americans played an integral role in the fire regime, accidentally and/or intentionally setting fire to prairie ecosystems (Day 1953, Chapman 1984). Where large-scale herbivores are abundant, grazing may help inhibit the succession of grass-dominated oak barrens to woodland (Ritchie et al. 1998).

The character of oak barrens can differ dramatically, primarily as the result of varying fire intensity and frequency, which are influenced by climatic conditions, soil texture, topography, and landscape context (i.e., proximity to water bodies and fire-resistant and fire-conducing plant communities) (Bowles et al. 1994, Chapman et al. 1995). Infrequent, high-intensity fires kill mature oaks and produce barrens covered by abundant scruffy oak sprouts. Park-like barrens with widely spaced trees and an open grass understory are maintained by frequent low-intensity fires, which occur often enough to restrict maturation of oak seedlings (Chapman et al. 1995).
Vegetation Description: The oak barrens community is a heterogeneous savanna vegetation type with variable physiognomy in time and space. Structurally, oak barrens range from dense thickets of brush and understory scrub oak within a matrix of grassland to park-like open woods of widely spaced mature oak with virtually no shrub or sub-canopy layer above the open forb and graminoid understory (Bowles and McBride 1994, Chapman et al. 1995, Kost et al. 2007). The physiognomic variations, which occur along a continuum, are the function of the complex interplay between fire frequency and intensity (Chapman et al. 1995). Typically, oak barrens grade into prairie on one edge and dry forest on the other. As noted by Bray (1958) and Curtis (1959), the flora of this community is a mixture of prairie and forest species, with prairie forbs and grasses more abundant in high light areas and forest forbs and woody species in the areas of low light.

The canopy layer generally varies from 5 to 60 percent cover (Chapman et al. 1989) and is dominated or co-dominated by *Quercus velutina* (black oak) and *Quercus alba* (white oak). These species of oak are also prevalent in the sub-canopy in shrubby clumps, especially where fire intensity is high. In addition, *Acer rubrum* (red maple), *Prunus serotina* (black cherry), *Populus grandidentata* (bigtooth aspen), *Populus tremuloides* (trembling aspen), and *Quercus ellipsoidalis* (pin oak) are often found in the overstory and sub-canopy of this community. Pin oak is especially common on excessively well-drained sites. Prevalent species of the subcanopy layer include: *Carya* spp. (hickory species), *Cornus* spp. (dogwood species), *Corylus americana* (American hazelnut), *Prunus* spp. (cherry species), and *Sassafras albidum* (sassafras).


In the absence of fire and with the prevalence of anthropogenic disturbance such as logging, off-road vehicle recreation, and livestock grazing, the following exotic species may be dominant components of the herbaceous layer of oak barrens: Agropyron repens (quack grass), Agrostis stolonifera (creeping bent), Asparagus officinalis (wild asparagus), Centaurea maculosa (spotted knapweed), Hieracium spp. (hawkweeds), Hypericum perforatum (St. John’s-wort), Poa compressa (Canada bluegrass), Poa pratensis (Kentucky bluegrass), Rumex acetosella (sheep sorrel), and Tragopogon dubius (goat’s beard).

**Michigan indicator species:**

**Spring/Early Summer**

Comandra umbellata (bastard toadflax), Coreopsis lanceolata (lanceolate coreopsis), Geum triflorum (prairie smoke, state threatened), Lithospermum canescens (hoary poucneas), Lupinus perennis, Krigia biflora, Pedicularis canadensis, Potentilla simplex (common cinquefoil), Senecio plattensis (prairie ragwort), Stipa spartea, and Viola pedata.

**Summer**

Anemone cylindrica (thimbleweed), Asclepias tuberosa (butterfly weed), Asclepias verticillata (whorled milkweed), Ceanothus americanus, Helianthus occidentalis (woodland sunflower), Helianthus divaricatus, Linum sulcatum (furrowed flax, state special concern), Monarda punctata (horsemint), Monarda fistulosa, Opuntia humifusa (prickly pear), and Trichostema dichotomum (bastard pennyroyal, state threatened).

**Fall**

Andropogon gerardii, Aristida purpurascens (three awn grass), Aster oolentangiensis, Aster ericoides (many flowered aster), Aster sericeus (silky aster, state threatened), Aureolaria flavia (false foxglove), Aureolaria pedicularia (false foxglove), Aureolaria virginica (false foxglove), Bouteloua curtipendula (side oats grama, state endangered), Liatris aspera (rough blazing star), Liatris cylindracea (dwarf blazing star), Schizachyrium scoparium, Silphium terebinthinaceum (prairie dock), Solidago speciosa (showy goldenrod), Solidago rigidia (stiff goldenrod), and Stipa avenacea.

**Other noteworthy species:** Rare plants associated with oak barrens include: Aristida tuberculosa (beach three-awned grass, state endangered), Aster drummondii (Drummond’s aster, state threatened), Astrer sericeus (western silvery aster, state threatened), Astragalus canadensis (Canadian milk-vetch, state threatened), Baptisia leucophaea (cream wild indigo, state endangered), Bouteloua curtipendula (side-oats grama grass, state endangered), Carex inops ssp. heliophila (sun sedge, state special concern), C. richardsonii (Richardson’s sedge, state special concern), C. tincata (tinged sedge, state threatened), Celtis tenuifolia (dwarf hackberry, state special concern), Cercastum velutinum (filed chickweed, presumed extirpated from Michigan), Cirsiium hillii (Hill's thistle, state special concern), Dasistoma macrophylla (mullein-foxglove, state endangered), Eupatorium sessilifolium (upland boneseed, state threatened), Euphorbia commutata (tinted spurge, state threatened), Gentiana puberulenta (downy gentian, state endangered), Geum triflorum (prairie-smoke, state threatened), Gillionia trifoliata (Bowman’s root, state endangered), Helianthus hirsutus (whiskered sunflower, state special concern), Helianthus mollis (downy sunflower, state threatened), Hieracium paniculatum (panicled hawkweed, state threatened), Houstonia caerulea (azure bluet, presumed extirpated from Michigan), Kuhnia eupatorioides (false boneseed, state special concern), Lechea minor (least pinweed, presumed extirpated from Michigan), Lechea stricta (erect pinweed, state special concern), Linum sulcatum (furrowed flax, state special concern), Paronychia fastigiata (low-forked chickweed, presumed extirpated from Michigan), Phlox bifida (cleft phlox, presumed extirpated from Michigan), Prunus alleghaniensis var. davisii (Alleghany or sloe plum, state special concern), Silene stellata (starry campion, state threatened), Silene virginica (fire pink, state endangered), Sisyrinchium strictum (blue-eyed grass, state special concern), Solidago bicolor (white goldenrod, state endangered), Sporobolus clandestinus (dropsseed, state endangered), and Trichostema dichotomum (bastard pennyroyal, state threatened).

Oak barrens frequently support numerous lichens and mosses. Oak trunks provide substrate for foliose lichens (e.g., Punctelia rudecta, Physcia millegrana, and Candelariella concolor) and crustose lichens (e.g., Candelariella xanthostigma and Rinodina papillata). In addition to the cryptogamous communities thriving on tree boles, oak barrens often contain patches of microbiotic soil crust composed of lichens, mosses, and cyanobacteria (Will-Wolf and Stearns 1999).

The oak barrens and surrounding prairie habitat share a rich diversity of invertebrates including numerous...
butterflies, skippers, grasshoppers, and locusts. However, the fragmented and degraded status of midwestern oak barrens/savannas and prairies has resulted in the drastic decline of numerous insect species associated with dry, open habitats or obligates of barrens and prairie host plants (Chapman et al. 1995). Rare butterflies, skippers, and moths include: Atrytonopsis hianna (dusted skipper, state special concern), Catocala amestris (three-staff underwing, state endangered), Erynnis p. persius (persius dusky wing, state threatened), Hesperia ottoe (ottoe skipper, state threatened), Incisalia henrici (Henry’s elfin, state threatened), Incisalia irus (frosted elfin, state threatened), Lycaeides melissa samuelis (Karner blue, state threatened/federal endangered), Papaipema sciuata (Culver’s root borer, state special concern), Pyrgactia spraguei (Sprague’s pygargtia, state special concern), Pyrgus centaureae wyandot (grizzled skipper, state special concern), Schinia indiana (phlox moth, state endangered), Schinia lucens (leadplant moth, state endangered), Spartiniphaga inops (Spartina moth, state special concern), and Speyeria idalia (regal fritillary, state endangered).

Other rare invertebrates include Lepyronia gibbosa (Great Plains spittlebug, state special concern), Oecanthus pini (pinetree cricket, state special concern), Orphulella pelidna (green desert grasshopper, state special concern), Prosapia ignipeptus (red-legged spittlebug, state special concern), and Scudderia fasciata (pine katydid, state special concern).

Numerous songbirds utilize oak barrens. Rare species include Ammodramus savannarum (grasshopper sparrow, state special concern) and Dendroica discolor (prairie warbler, state endangered). Typical songbirds include: Melospiza lincolni (Lincoln’s sparrow), Passerina cyanea (indigo bunting), Pooecetes gramineus (vesper sparrow), Sial sialis (eastern bluebird), Spizella passerina (chipping sparrow), Spizella pusilla (field sparrow), Toxostoma rufum (brown thrasher), Vermivora pinus (blue-winged warbler) and Vermivora ruficapilla (Nashville warbler). Additional avian species that utilize this habitat include: Accipiter striatus (sharp-shinned hawk), Bartamia longicauda (upland sandpiper), Bonasa umbellus (ruffed grouse), Buteo jamaicensis (red-tailed hawk), Carduelis tristis (American goldfinch), Charadrius vociferus (killdeer), Chondestes grammacus (lark sparrow), Falco sparverius (American kestrel), Icterus galbula (Baltimore oriole), Melanerpes erithrocephalus (red-headed woodpecker), Meloagris gallopavo (wild turkey), Otus asio (Eastern screech-owl), Tyrannus tyrannus (eastern kingbird), and Zenaida macroura (mourning dove). Savanna restoration with prescribed fire in Minnesota resulted in the increase in the abundance of open-country bird species, including many species that have been declining in central and eastern North America (Davis et al. 2000).

Cryptotis parva (least shrew, state threatened) and Microtus ochrogaster (prairie vole, state endangered) are rare mammals that may be found in oak-pine barrens. Additional mammals commonly associated with the oak barrens community include: Canis latrans (coyote), Microtus pennsylvanicus (meadow vole), Odocoileus virginianus (white-tailed deer), Scirius niger (fox squirrel), Spermophilus tridecemlineatus (thirteen-lined ground squirrel), Taxidea taxus (badger), Vulpes vulpes (red fox), and Zapus hudsonia (jumping meadow mouse).

Several rare reptiles are known from this community type. They include: Pantherophis spiloides (gray ratsnake, state special concern), Sistrurus c. catenatus (eastern massasauga, state special concern, federal candidate species), and Terrapene c. carolina (eastern box turtle, state special concern). Some of the more common amphibians and reptiles that frequent the oak barrens include: Bufo a. americanus (eastern American toad), Bufo fowleri (Fowler’s toad), Heterodon platirhinos (eastern hog-nosed snake), and Opheodrys vernalis (smooth green snake).

Conservation/management: Fire is the single most significant factor in preserving the oak barrens landscapes. Where remnants of oak barrens persist, the use of prescribed fire is an imperative management tool for maintaining an open canopy, promoting high levels of grass and forb diversity, deterring the encroachment of woody vegetation and invasive exotics, and limiting the success of dominants. Numerous studies have indicated that fire intervals of 1-3 years bolster graminoid dominance, increase overall grass and forb diversity, and remove woody cover of saplings and shrubs (White 1983, Tester 1989). Burning at longer time intervals will allow for seedling establishment and the persistence of woody plants. Where rare invertebrates are a management concern, burning strategies should allow for ample refugia to facilitate effective postburn recolonization (Michigan Natural Features Inventory 1995, Siemann et al. 1997).

Though most of the historical oak barrens have been degraded by selective logging, livestock grazing, and fire
suppression or destroyed by development, agricultural clearing, sand mining, and extensive timber harvest, there is much opportunity for restoration of this community type. Plant species of oak barrens can persist through cycles of canopy closure and removal (Chapman et al. 1995). The occurrence of oak barrens indicator species in closed-canopy forests reveals the presence of a native seedbank and highlights that area as a target for restorative management. Also indicative of a site’s potential for restoration is the prevalence of oak “wolf trees.” “Wolf trees” are large open-grown trees with wide-spreading limbs that are often associated with oak barrens’ plants or seedbank (Michigan Natural Features Inventory 1995).

Where canopy closure has degraded the savanna character, one can restore the oak barrens community by selectively cutting the majority of trees (White 1986), leaving an average of 4 trees/acre. Degraded barrens that have been long deprived of fire often contain a heavy overstory component of shade tolerant species, which can be removed by mechanical thinning (Peterson and Reich 2001). Reconstructed sites will need to be maintained by periodic prescribed fire and may require investment in native plant seeding where seed and plant banks are inadequate (Packard 1988). Depending on the physiognomic target of the management, one can manipulate the intensity, seasonality, and frequency of the prescribed burns: low-intensity and high-frequency burns for the park-like end of the barrens continuum and low-frequency and high-intensity burns for shrubby oak barrens. Fall burns typically are slow moving, low-intensity fires due to high relative humidity and slow wind speed, while late spring and summer burns are often more intense due to higher wind speeds and lower relative humidity (King 2000). Summer burning can be employed to simulate naturally occurring lightning season burns. Early spring burns often carry irregularly through barrens influenced by high spring water tables. Such patchy burns can be useful tool for establishing refugia for fire-sensitive species and may permit oak seedling establishment (Chapman et al. 1995). Patchy burns are often the result of frequent low-intensity fires, which carry sporadically through areas with low fuel loads. In contrast, infrequent fires are often more uniform in coverage, spreading evenly through areas of high fuel accumulation (Ladd 1991).

In many circumstances the effective use of prescribed fire is precluded due to monetary constraints or safety concerns. In areas where fire is undesirable or unfeasible, mowing or selective cutting can be utilized and should be carried forth in late fall or winter to minimize detrimental impact to herbaceous species and rare invertebrates (Chapman et al. 1995, Michigan Natural Features Inventory 1995, King 2000). Management of oak barrens communities should be orchestrated in conjunction with the management of adjacent communities such as dry sand prairie, dry southern forest, and coastal plain marsh.

Research needs: As noted by Nuzzo (1986), Minc and Albert (1990), Faber-Langendoen (1993), and Bowles and McBride (1994), no single definition of Midwest oak savanna or oak barrens is universally accepted, and numerous distinct community types have been lumped under the phrase “Midwest oak savanna.” Misunderstanding and misuse of the term can be alleviated by the continued refinement of regional classifications that correlate species composition, site productivity, ecological process, and landscape context. Understanding spatial and temporal variability of oak barrens is also crucial for determining the direction of management.

Management of oak barrens remnants can be determined by site-specific research of site characteristics and presettlement composition and structure (Minc and Albert 1990, Bowles et al. 1994, Bowles and McBride 1998). Investigation into the frequency, periodicity (seasonality), and intensity of fires in oak barrens is needed to guide restoration and management activities. In addition, because limitations imposed by safety concerns can hamper the effectiveness of prescribed fire as a management tool, maintaining the ecological integrity of oak barrens requires experimentation with different disturbance combinations (King 2000). Effects of management need to be monitored to allow for assessment and refinement.
Since all of Michigan’s oak barrens are degraded, it is essential to determine what role seedbanks, vegetative reproduction, and external seed sources play in restoration of remnant barrens. Numerous rare Lepidoptera have host plants occurring on oak barrens. The effects of fire and alternative management techniques on rare faunal populations and their host vegetation need to be studied (Chapman et al. 1995, Siemann et al. 1997). In addition, because of the daunting problem of exotic species encroachment, research needs to examine management strategies that minimize invasive species introduction and dominance.

**Similar communities:** Bur oak plains, dry sand prairie, dry southern forest, dry-mesic prairie, lakeplain oak openings, oak openings, oak-pine barrens, and pine barrens.

**Other Classifications:**

**Michigan Natural Features Inventory Presettlement Vegetation (MNFI):**
Black Oak Barren and Mixed Oak Savanna

**Michigan Department of Natural Resources (MDNR):** G-grass and O0(zero)-oak with <100 trees per acre.

**Michigan Resource Information Systems (MIRIS):**
33 (Pine or Oak Opening), 412 (Central Hardwood), 4122 (White Oak), 4123 (Black Oak), 4129 (Other Oak).

**The Nature Conservancy National Classification:**

**CODE; ALLIANCE; ASSOCIATION; COMMON NAME**

II.B.2.N.a.12; *Quercus alba* - (*Quercus velutina*) Woodland Alliance; *Quercus velutina* - (*Quercus ellipsoidalis*) - *Quercus alba/Deschampsia flexuosa* Woodland; Black Oak- Northern Pin Oak / Common Hairgrass Woodland

V.A.6.N.e.3; *Quercus velutina* - (*Quercus ellipsoidalis*) Wooded Herbaceous Alliance; *Quercus velutina* - (*Quercus alba*) - *Quercus ellipsoidalis/Secale montana scoparius* - *Lupinus perennis* Wooded Herbaceous Vegetation; Black Oak/Lupine Barrens

**Related Abstracts:** Alleghany plum, bur oak plains, Culver’s root borer, dry sand prairie, dry-mesic prairie, Hill’s thistle, karner blue butterfly, lakeplain oak openings, oak barrens, oak openings, oak-pine barrens, pine barrens, prairie smoke, and red-legged spittlebug.

**Selected References:**


Michigan Natural Features Inventory. 1995. Forest stewardship training materials for oak-pine barrens ecosystem. (Unpublished manuscript.) Michigan Natural Features Inventory, Lansing, MI.


Abstract Citation:

Updated June 2010.

Copyright 2004 Michigan State University Board of Trustees.

Michigan State University Extension is an affirmative-action, equal-opportunity organization.

Funding for abstract provided by Michigan Department of Natural Resources- Forest Management Division and Wildlife Division.

Recently burned oak barrens remnant, Fort Custer, Calhoun County. Photo by Joshua G. Cohen.