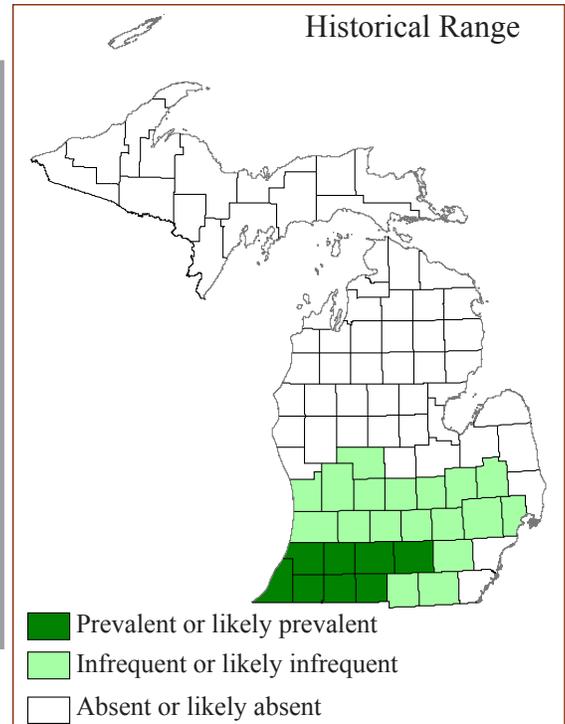




Photo by Bradford S. Slaughter



Overview: Dry-mesic prairie is a native grassland community dominated by big bluestem (*Andropogon gerardii*), little bluestem (*Andropogon scoparius*), and Indian grass (*Sorghastrum nutans*) that occurs on sandy loam or loamy sand on level to slightly sloping sites of glacial outwash, coarse-textured end moraines, and glacial till plain. The community represents the stands of open grassland that occurred within the historic oak openings. Areas dominated by native grasses with less than one mature tree per acre (0.4 ha) are considered prairie (Curtis 1959). This natural community type was known as woodland prairie in previous versions of the natural community classification (see Kost et al. 2007).

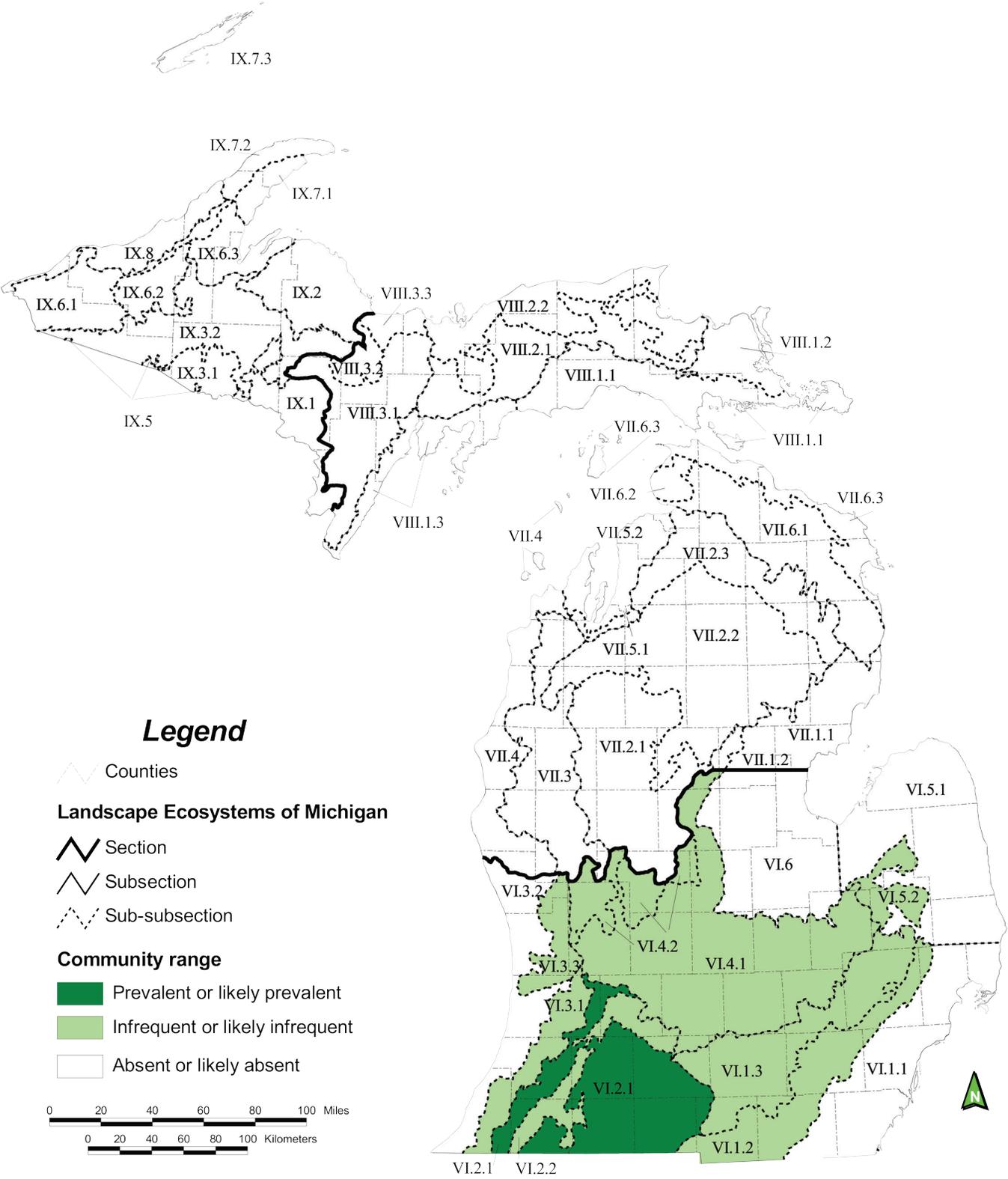
Global and State Rank: G3/S1

Range: In the 1800s, dry-mesic prairie occurred in association with oak openings throughout much of southern Lower Michigan. Based on interpretations of General Land Office surveyor notes from the early to mid 1800s, pockets of dry-mesic prairie occurred in Barry, Berrien, Branch, Calhoun, Cass, Ionia, Kalamazoo, Livingston, Lapeer, St. Joseph, Van Buren, and Washtenaw counties (Comer et al. 1995). Oak openings, which likely included pockets of dry-mesic prairie, occurred throughout the upper Midwest and stretched into western New York and southern Ontario, Canada (Faber-Langendoen 2001).

Rank Justification: In the early to mid 1800s, the southern Lower Peninsula supported approximately 60,500 acres (24,500 ha) (Comer et al. 1995) of upland prairie, which included pockets of dry-mesic prairie, mesic prairie, mesic sand prairie, dry sand prairie, and hillside prairie. Because of its close association with oak openings, dry-mesic prairie represented the most widespread type of prairie community in southern Michigan. The Michigan Natural Features Inventory database currently includes 11 occurrences of dry-mesic prairie, which range in size from 2 to 15 acres (ave. 5.6 acres, or 2.3 ha) and total 62 acres (25 ha). It is difficult to reliably determine the total acreage of dry-mesic prairie in Michigan in the 1800s. However, based on comparisons of the total acreage of all upland prairie element occurrences in southern Lower Michigan today (480 acres, 194 ha) with that found in the early to mid 1800s (provided above), it appears that less than 1% of the original upland prairie remains intact.

Landscape and Abiotic Context: Dry-mesic prairie occurs primarily on level to slightly sloping sites of glacial outwash or coarse-textured end moraines on glacial outwash (Chapman 1984). Soils are typically sandy loam or occasionally loamy sand with pH ranging from 5.2 to 6.7 (ave. pH 5.8) and water retaining capacity of 43 to 94% (ave. 55%) (Chapman 1984). The majority of historical dry-mesic prairies occur within the Kalamazoo





Ecoregional map of Michigan (Albert 1995) depicting historical distribution of dry-mesic prairie (Albert et al. 2008)



Interlobate Subsection (Albert 1995), which represents the northernmost portion of the “Prairie Peninsula” described by Transeau (1935).

Dry-mesic prairie occurred historically within oak openings and may have graded into mesic prairie and bur oak plains on level outwash plains like the Battlecreek Outwash Plain (Albert 1995). Today the community is almost entirely restricted to railroad right-of-ways, which often border agricultural fields.

Natural Processes: Fire played a critical role in maintaining open conditions in Michigan prairie and oak savanna ecosystems. In the absence of frequent fires, which retarded woody growth prior to Euro-American settlement, Michigan’s prairies and open oak ecosystems (e.g., oak openings, bur oak plains, oak barrens, oak woodlands) were quickly colonized by trees and shrubs and converted to oak forests.

In 1835 Hoffman recounts his impression of a fire in December “To-day, for the first time, I saw the meadows on fire. They are of vast extent, running far into the woods like the friths of a lake; and as wild grass, which they supply in the greatest profusion, furnishes the new settler with all the hay he uses for his stock, they are burnt over thus annually to make it tender. These fires traveling far over the country seize upon the largest prairies, and consuming every tree in the woods, except the hardest, cause the often-mentioned oak openings, so characteristic of Michigan scenery. It is a beautiful sight to see the fire shooting in every direction over these broad expanses of land...” (Hoffman 1835 in Chapman 1984).

Van Buren in 1884 on describing the oak openings of Calhoun County writes “The annual fires burnt up the underwood, decayed trees, vegetation, and debris, in the oak openings, leaving them clear of obstructions. You could see through the trees in any direction, save where the irregularity of the surface intervened, for miles around you, and you could walk, ride on horse-back, or drive in a wagon wherever you pleased in this woods, as freely as you could in a neat and beautiful park.” (Van Buren 1884 in Chapman 1984).

Many early accounts have been written of the rapid conversion of prairie and oak savanna (oak barrens and oak openings) to forest. Glidden in 1892 describes the origin of many of the oak forests now occurring in southern Michigan in the following passage. “After the very best

job of breaking [plowing], a live [oak] grub would be left upon every square rod of ground. There is nothing now to compare with this pioneer grub. For fifty years or more its yearly growth had been burned off, and had sprouted again in the spring... The enlargement at the surface about the tap root increased with each year’s growth of sprouts, until the cap was formed, a foot or more in with, like an underground toad stool, although not so regular in shape. The whole under-surface of this cap was filled with dormant buds, that awoke in activity at once when the standing ones were cut or were burned away. Nature reasserted itself when the annual burnings had ceased, and the fittest stem survived and became the tree or young oak, as we see them today, while the cap has rotted away” (Glidden in Chapman 1984).

The conversion of open prairie and oak opening to forest was well underway by 1872 when Hubbard described the landscape near Pontiac as it appeared during his 1837 expedition with Douglas Houghton through Michigan’s “new territory”. Hubbard writes “I speak in the past tense, because, though the rural beauty of the country is still unrivaled, little remains of the original character of the [oak] openings. This is a result partly of the progress of cultivation, and partly of the thick growth of small timber that has covered all the uncultivated portions since the annual fires have ceased, which kept down the underbrush” (Hubbard 1872 in Chapman 1984).

Fire frequency depended on a variety of factors including type and volume of fuel, topography, natural firebreaks, and density of Native Americans (Chapman 1984). In general, the probability for a wide-ranging fire increases in level topography like large outwash plains (Chapman 1984). While occasional lightning strikes resulted in fires that spread across the landscape, Native Americans were the main sources of ignition.

There are many early accounts of Native Americans intentionally setting fires to accomplish specific objectives (see Day 1953, Curtis 1959, Thompson and Smith 1970, Chapman 1984, Denevan 1992, Kay 1995). Native Americans intentionally set fires in the fall to clear briars and brush and make the land more easily passable. Frequent fires kept the land open, increasing both short- and long-range visibility, which facilitated large game hunting and provided a measure of safety from surprise attacks by neighboring tribes. Fire was used to increase productivity of berry crops and agricultural fields. As a habitat management tool, fires were used to maintain high quality forage



for deer, elk, woodland caribou, bison and other game species. It was also used as a hunting tool to both drive and encircle game. During warfare, fire was strategically employed to drive away advancing enemies, create cover for escape, and for waging attacks.

In addition to maintaining open conditions, fire also plays a critical role in maintaining species diversity. A recensus of 54 prairie remnants in Wisconsin found that 8 to 60% of the original plant species recorded at the sites had been lost over time (32 to 52 years) even though the sites appeared relatively undisturbed (Leach and Givnish 1996). The authors suggest that taller vegetation outcompeted species with small statures, small seeds (e.g., orchids), and nitrogen-fixing symbioses such as members of the legume family (Fabaceae) like lupine (*Lupinus perennis*), wild indigo (*Baptisia* spp.), bush clover (*Lespedeza* spp.), and tick-trefoil (*Desmodium* spp.). Because fire maintains open conditions and burns off standing and accumulated litter, small species and those with small seeds that require open microsites are able to garner enough space and light to remain viable. In the absence of frequent fires, small species are outcompeted by taller and denser types of vegetation. As fire volatilizes much of the nitrogen stored in combustible vegetation, frequent burning also favors species that form nitrogen-fixing symbioses (e.g., legumes and rhizobium bacteria) and thus provides these plants with a competitive edge not found in unburned sites (Leach and Givnish 1996).

Fire also helps maintain species diversity by facilitating expression of the soil seed bank and promoting seed germination and establishment. By consuming accumulated and standing leaf litter, fire increases light availability to the soil surface and increases diurnal temperature fluctuations, both of which trigger seed germination. In addition, the removal of litter by fire creates critical microsites for seed germination and fosters seedling establishment.

The removal of litter by fire also increases the availability of many important plant nutrients (e.g., N, P, K, Ca and Mg), which are thought to contribute to higher plant biomass, increased flowering and seed production, and greater palatability to herbivores (Vogl 1964, Daubenmire 1968, Viro 1974, Vogl 1974, Smith and Kadlec 1985, Abrams et al. 1986, Collins and Gibson 1990, Reich et al. 1990, Schmalzer and Hinkle 1992, Timmins 1992, Laubhan 1995, Warners 1997).

While this discussion has focused on plants it is important to note that these species serve as host plants for a variety of insects and the structure of open grasslands is critical to a wide variety of animal species, many of which are considered rare or declining today.

Ants, particularly the genus *Formica*, play an important role in mixing and aerating prairie soils (Curtis 1959, Trager 1998). Large ant mounds, which may measure .5 m in height and over 1 m wide and number 40 to 50 per acre are especially conspicuous following a prairie fire (Curtis 1959). Because of their abundance and frequent habit of abandoning old mounds and building new ones, ants overturn large portions of prairies in a relatively short time (Curtis 1959). Other important species contributing to soil mixing and aeration include moles, mice, skunks, and badgers (Curtis 1959).

Historically, large herbivores such as bison significantly influenced plant species diversity in Michigan prairie and oak savanna ecosystems. The diet of bison consists of 90 to 95% grasses and sedges (Steuter 1997). As bison selectively forage on grasses and sedges, they reduce the dominance of graminoids and provide a competitive advantage to forb species. The activities of bison, which includes wallowing and trampling, promotes plant species diversity by creating microsites for seed germination and seedling establishment and reducing the dominance of robust perennials (Steuter 1997).

Vegetation Description: Unfortunately, no detailed ecological study of dry-mesic prairie was completed in Michigan before the nearly total demise of the community. What information is available comes from written descriptions of oak openings by early European settlers and from studies of small prairie remnants.

Chapman (1984) completed a study of 66 prairie and savanna remnants in southern Lower Michigan, thirteen of which he classified as dry-mesic prairie. In addition, Curtis (1959) and Curtis and Green (1949) collected detailed information on 66 dry-mesic prairie stands in Wisconsin and much of their data may be applicable to dry-mesic prairie in Michigan.

Dry-mesic prairie supports a dense to moderately dense growth of low to medium vegetation with very little bare ground (Chapman 1984). The community is dominated by big bluestem, little bluestem, and Indian grass, which can occur in varying degrees of dominance to one another



(Chapman 1984). Switch grass (*Panicum virgatum*) occurred in only three of the thirteen dry-mesic prairie sites studied by Chapman and was not listed among the prevalent species for dry-mesic prairie in Wisconsin by Curtis (1959). Species that reach their greatest abundance (e.g., modal species) in dry-mesic prairie in Michigan include the following: lead plant (*Amorpha canescens*, state special concern), thimbleweed (*Anemone cylindrica*), butterfly weed (*Asclepias tuberosa*), smooth aster (*Aster laevis*), and daisy fleabane (*Erigeron strigosus*) (Chapman 1984).

Oak grubs of white oak, black oak, and bur oak, which were maintained in a shrub-like condition as a result of annual fires, were abundant in dry-mesic prairie as were widely scattered, open grown adults of these same species, especially white oak.

The following table of dry-mesic prairie plants was compiled from Chapman's (1984) study of thirteen dry-mesic prairie remnants in southern Lower Michigan and includes only species occurring in more than half the sites he studied.

SCIENTIFIC NAME	COMMON NAME
Grasses and Sedges	
<i>Andropogon gerardii</i>	big bluestem
<i>Andropogon scoparius</i>	little bluestem grass
<i>Carex pensylvanica</i>	Pennsylvania sedge
<i>Sorghastrum nutans</i>	Indian grass
Forbs	
<i>Achillea millefolium</i>	yarrow
<i>Anemone cylindrica</i>	thimbleweed
<i>Antennaria parlinii</i>	smooth pussytoes
<i>Asclepias syriaca</i>	common milkweed
<i>Asclepias tuberosa</i>	butterfly weed
<i>Aster laevis</i>	smooth aster
<i>Aster oolentangiensis</i>	prairie heart-leaved aster
<i>Aster pilosus</i>	hairy aster
<i>Comandra umbellata</i>	bastard toadflax
<i>Desmodium canadense</i>	showy tick-trefoil
<i>Desmodium illinoense</i>	prairie tick-trefoil
<i>Desmodium marilandicum</i>	small-leaved tick-trefoil
<i>Erigeron strigosus</i>	daisy fleabane
<i>Euphorbia corollata</i>	flowering spurge
<i>Fragaria virginiana</i>	wild strawberry
<i>Helianthus occidentalis</i>	western sunflower
<i>Hieracium longipilum</i>	long-bearded hawkweed
<i>Lactuca canadensis</i>	tall lettuce
<i>Lespedeza capitata</i>	round-headed bush-clover

<i>Lithospermum canescens</i>	hoary puccoon
<i>Monarda fistulosa</i>	wild bergamot
<i>Potentilla simplex</i>	old field cinquefoil
<i>Ratibida pinnata</i>	yellow coneflower
<i>Rudbeckia hirta</i>	black-eyed susan
<i>Smilacina racemosa</i>	false spikenard
<i>Solidago juncea</i>	early goldenrod
<i>Solidago nemoralis</i>	old field goldenrod
<i>Solidago rigida</i>	stiff goldenrod
<i>Solidago speciosa</i>	showy goldenrod
<i>Tradescantia ohiensis</i>	common spiderwort

Shrubs

<i>Ceanothus americanus</i>	New Jersey tea
<i>Rosa carolina</i>	pasture rose
<i>Rubus flagellaris</i>	northern dewberry
<i>Salix humilis</i>	prairie willow

Michigan Indicator Species: Chapman lists only one species, hairy aster (*Aster pilosus*), as an indicator for dry-mesic prairie in Michigan. Thimbleweed (*Anemone cylindrica*) and western sunflower (*Helianthus occidentalis*), which are listed as indicators of dry-mesic prairie in Wisconsin by Curtis (1959) and are common in Michigan dry-mesic prairies, may be suitable indicator species for dry-mesic prairie in Michigan.

Other Noteworthy Species: Rare plant species associated with dry-mesic prairie are listed below along with their status, which is indicated by the following abbreviations: X, extirpated from state; E, State Endangered; T, State Threatened; SC, State Species of Special Concern; LE, Federally Endangered.

Scientific Name	Common Name	Status
<i>Amorpha canescens</i>	lead-plant	SC
<i>Baptisia lactea</i>	white false indigo	SC
<i>Carex inops</i> ssp. <i>heliophila</i>	sun sedge	SC
<i>Cirsium hillii</i>	Hill's thistle	SC
<i>Coreopsis palmata</i>	prairie coreopsis	T
<i>Echinacea purpurea</i>	purple coneflower	X
<i>Gentiana flavida</i>	white gentian	E
<i>Panicum leibergii</i>	Leiberg's panic grass	T
<i>Scleria triglomerata</i>	tall nut-rush	SC
<i>Viola pedatifida</i>	prairie birdfoot violet	T

Rare animal species associated with dry-mesic prairie include the following:



Grassland Birds: Henslow's sparrow (*Ammodramus henslowii*) (E), grasshopper sparrow (*Ammodramus savaannarum*) (SC), short-eared owl (*Asio flammeus*) (E), long-eared owl (*Asio otus*) (T), northern harrier (*Circus cyaneus*) (SC), migrant loggerhead shrike (*Lanius ludovicianus migrans*) (E), dickcissel (*Spiza americana*) (SC), western meadowlark (*Sturnella neglecta*) (SC), and barn owl (*Tyto alba*) (E).

Insects: blazing star borer (*Papaipema beeriana*) (SC), phlox moth (*Schinia indiana*) (E), leadplant flower moth (*Schinia lucens*) (E), red-legged spittlebug (*Prosapia ignipectus*) (SC), Sprague's pygarcia (*Pygarcia spraguei*) (SC), American burying beetle (*Nicrophorus americanus*) (X/LE), pinetree cricket (*Oecanthus pini*) (SC), and regal fritillary (*Speyeria idalia*) (E).

Mammals: prairie vole (*Microtus ochrogaster*) (E).

Reptiles: eastern massasauga (*Sistrurus c. catenatus*) (SC and Federal Candidate Species), gray ratsnake (*Pantherophis spiloides*) (SC), and eastern box turtle (*Terrapene c. carolina*) (SC). Spotted turtle (*Clemmys guttata*) (T) and Blanding's turtle (*Emydoidea blandingii*) (SC) may nest in dry-mesic prairie when it occurs adjacent to wetlands.

Conservation and Biodiversity Management: Efforts should be made to identify, protect, and manage remnants of dry-mesic prairie where they occur. Several studies to identify prairie remnants in Michigan have been undertaken and most remnants are very small and/or occur as narrow strips adjacent to railroads (Scharrer 1972, Thompson 1970, 1975 and 1983, Chapman 1984). The small size and poor landscape context of most remnant dry-mesic prairies makes large-scale restoration of existing prairies nearly impossible. Prairie plantings located in areas of former dry-mesic prairie in southwestern Lower Michigan are particularly needed.

Managing dry-mesic prairie requires frequent burning, from annual to every two to three years. Longer burn intervals will result in tree and tall shrub encroachment. Prescribed burning is required to protect and enhance plant species diversity and prevent encroachment of trees and tall shrubs, which outcompete light-demanding prairie plants. In prairie remnants where fire has been excluded for long periods (e.g., decades), local extinctions of plant species are common (Leach and Givnish 1996).

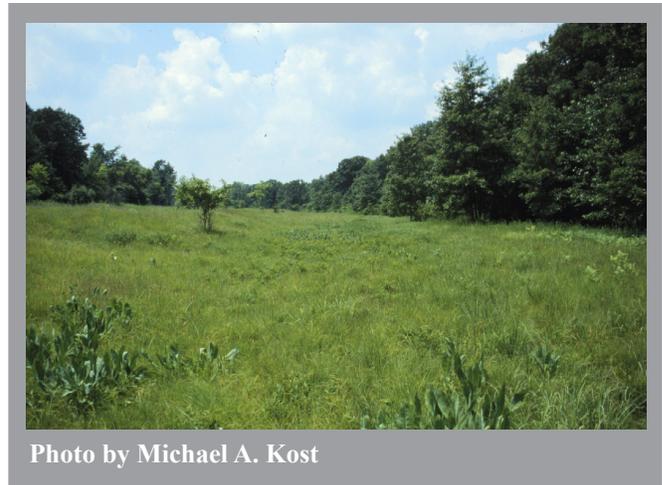


Photo by Michael A. Kost

Dry-mesic prairie remnants in Michigan are largely restricted to railroad rights-of-way.

In addition to prescribed fire, brush cutting accompanied by herbicide application to cut stumps is an important component of prairie restoration. While fires frequently kill woody seedlings, long established trees and tall shrubs like black cherry (*Prunus serotina*) and dogwoods (*Cornus* spp.) typically resprout and can reach former levels of dominance within two to three years. Herbicide application to cut stumps will prevent resprouting.

To reduce the impacts of management on fire-intolerant species it will be important to consider a rotating schedule of prescribed burning in which adjacent management units are burned in alternate years. This is especially important when planning burns in open grasslands such as dry-mesic prairie. Insect species that are restricted to these habitats have already experienced severe losses in the amount of available habitat due to forest succession brought on by years of fire suppression. By burning adjacent management units in alternate years, insect species from unburned units may be able to recolonize burned areas (Panzer et al. 1995). Avian species diversity is also thought to be enhanced by managing large areas as a mosaic of burned and unburned patches (Herkert et al. 1993).

Prairie ants (*Formica*) are an extremely important component of grassland communities and research indicates that they respond with population increases to restoration activities, especially prescribed fire (Trager 1998). Prescribed burning precipitates changes in the dominance of ant species from carpenter and woodland ants (*Camponotus* and *Aphaenogaster*) to prairie ants because it reduces woody vegetation and detritus used by the arboreal and litter- and twig-nesting species in favor of species restricted



to grassland habitats (Trager 1998). Restorations involving prairie plantings near old fields or remnant prairies are typically colonized by several species of prairie ants within a few years (Trager 1990).

Controlling invasive species is a critical step in restoring and managing dry-mesic prairie. By outcompeting native species, invasives alter vegetation structure, reduce species diversity, and upset delicately balanced ecological processes such as trophic relationships, interspecific competition, nutrient cycling, soil erosion, hydrologic balance, and solar insolation (Bratton 1982, Harty 1986). At present some of the most aggressive invasive species that threaten biodiversity of grassland communities include reed canary grass (*Phalaris arundinacea*), spotted knapweed (*Centaurea maculosa*), white and yellow sweet clover (*Melilotus alba* and *M. officinalis*), autumn olive (*Elaeagnus umbellata*), multiflora rose (*Rosa multiflora*), common buckthorn (*Rhamnus cathartica*), Eurasian honeysuckles (*Lonicera maaackii*, *L. morrowii*, *L. tatarica*, *L. x bella.*), and black locust (*Robinia pseudoacacia*).

In addition to reestablishing ecological processes such as prescribed fire, most restoration sites will require the reintroduction of appropriate native species and genotypes. Plants can be reintroduced through both seeding and seedling transplants. Small, isolated prairie remnants may harbor plant populations that have suffered from reduced gene flow. Restoration efforts at isolated prairie remnants should consider introducing seeds collected from nearby stocks to augment and maintain genetic diversity of remnant plant populations. The Michigan Native Plant Producers Association may be a helpful resource for locating sources of Michigan genotypes (<http://www.nohlc.org/MNPPA.htm>).

Several helpful guides are available for restoring prairies and starting prairie plants from seed (Packard and Mutel 1997, Nuzzo 1976, Schulenberg 1972). See Packard and Mutel (1997) for a comprehensive treatment of the subject and additional references.

Restoration and management of grasslands such as dry-mesic prairie are critically important to grassland birds, which have suffered precipitous population declines due to habitat loss and changing agricultural practices (e.g., early mowing of hay fields). Detailed habitat management guidelines for grassland birds have been developed by Herkert et al. (1993) and Sample and Mossman (1997). Listed below are several of the recommendations sug-

gested by Herkert et al. (1993) (see publication for complete list of management guidelines).

1. Avoid fragmentation of existing grasslands.
2. Grassland restorations aimed at supporting populations of the most area-sensitive grassland birds should be at least 125 acres and preferably more than 250 acres in size. Area sensitive species requiring large patches of grassland (>100 acres) include northern harrier (SC), bobolink (*Dolichonyx oryzivorus*), savannah sparrow (*Passerculus sandwichensis*), Henslow's sparrow (SC), grasshopper sparrow (SC), eastern meadowlark (*Sturnella magna*), western meadowlark (SC), sedge wren (*Cistothorus platensis*), sharp-tailed grouse (*Pedioecetes phasianellus*), upland sandpiper (*Bartramia longicauda*), short-eared owl (E), and barn owl (E) (Herkert et al. 1993, Sample and Mossman 1997). Patches of grassland less than 50 acres will benefit the least area-sensitive grassland birds such as northern bobwhite (*Colinus virginianus*), red-winged black bird (*Agelaius phoeniceus*), American goldfinch (*Carduelis tristis*), Vesper sparrow (*Poocetes gramineus*), field sparrow (*Spizella pusilla*), song sparrow (*Melospiza melodia*), dickcissel (SC), and common yellowthroat (*Geothlypis trichas*) (Herkert et al. 1993).
3. Maximize interior grassland habitat by establishing circular (best) or square grassland plantings and avoiding long, narrow plantings, which increase edge habitat.
4. Where grassland habitats border forests, strive to create a feathered edge by allowing prescribed fires to burn through adjacent forests as opposed to installing firebreaks along the forest edge. Grasslands with feathered edges experience lower rates of nest predation than those with sharply contrasting edges (Ratti and Reese 1988).



Research Needs: Remaining remnants of dry-mesic prairie need to be identified, protected, and managed. Further research on the historical plant species composition of dry-mesic prairie in Michigan would be useful for developing seed mixes for restoration. Genetic studies of the effects of small, isolated populations on plant species genetic diversity will provide information on managing remnants of dry-mesic prairie. Research on the utilization of restored and remnant prairies by grassland birds will provide useful information for understanding how dry-mesic prairies contribute to biodiversity. Studies on methods of prairie establishment and management, including controlling invasive species, will benefit both ongoing and new efforts to restore dry-mesic prairie. Conservation and management efforts will benefit from further study of how species composition is influenced by fire frequency, intensity, and periodicity.

Similar Communities: oak openings, dry sand prairie, hillside prairie, mesic sand prairie, oak barrens, bur oak plains, and mesic prairie.

Other Classifications: Michigan Natural Features Inventory Circa 1800s Vegetation (MNFI): Grassland.

Michigan Department of Natural Resources (MDNR): G

The Nature Conservancy U.S. National Vegetation Classification and International Classification of Ecological Communities (Faber-Langendoen 2001, NatureServe 2004): Michigan dry-mesic prairie is not recognized as a separate prairie type but is instead lumped with oak openings.

CODE; ALLIANCE; ASSOCIATION; COMMON NAME

V.A.6.N.c.2; *Quercus macrocarpa* – (*Quercus alba*) Wooded Herbaceous Alliance; *Quercus alba* – *Quercus macrocarpa* / *Andropogon gerardii* Wooded Herbaceous Vegetation; White Oak – Bur Oak Openings

Related Abstracts: oak openings, dry sand prairie, oak barrens, bur oak plains, mesic prairie, Culver's root borer, eastern box turtle, eastern massasauga, Henslow's sparrow, migrant loggerhead shrike, northern harrier, and red-legged spittlebug.

Literature Cited:

- Abrams, M.D., A.K. Knapp and L. C. Hulbert. 1986. A ten year record of aboveground biomass in a Kansas tallgrass prairie: Effects of fire and topographic position. *American Journal of Botany* 73:1509-15.
- Albert, D.A. 1995. Regional landscape ecosystems of Michigan, Minnesota, and Wisconsin: A working map and classification. Gen. Tech. Rep. NC-178. St. Paul, MN: USDA, Forest Service, North Central Forest Experiment Station, St. Paul, MN. <http://nrs.fs.fed.us/pubs/242> (Version 03JUN1998). 250 pp.
- Albert, D.A., J.G. Cohen, M.A. Kost, B.S. Slaughter, and H.D. Enander. 2008. Distribution maps of Michigan's Natural Communities. Michigan Natural Features Inventory, Report No. 2008-01, Lansing, MI. 174 pp.
- Bratton, S.P. 1982. The effects of exotic plant and animal species on nature preserves. *Natural Areas Journal*. 2(3):3-13.
- Chapman, K.A. 1984. An ecological investigation of native grassland in southern Lower Michigan. M.S. Thesis, Western Michigan University, Kalamazoo, MI. 235 pp.
- Collins, S.L. and D.J. Gibson. 1990. Effects of fire on community structure in tallgrass and mixed grass prairie. Pp. 81-98 in S. L. Collins and L. L. Wallace (eds.), *Fire in North American tallgrass prairies*, University of Oklahoma Press, Norman, OK.
- Comer, P.J., D.A. Albert, H.A. Wells, B.L. Hart, J.B. Raab, D.L. Price, D.M. Kashian, R.A. Corner and D.W. Schuen. 1995. Michigan's presettlement vegetation, as interpreted from the General Land Office Surveys 1816-1856. Michigan Natural Features Inventory, Lansing, MI. Digital Map.
- Curtis, J.T. 1959. *Vegetation of Wisconsin*. The University of Wisconsin Press, Madison, WI. 657 pp.
- Curtis, J.T. and H.C. Greene. 1949. A study of relic Wisconsin prairies by the species-presence method. *Ecology* 30:152-55.
- Daubenmire, R. 1968. Ecology of fire in grasslands. *Advances in Ecological Research* 5:209-66.
- Day, G.M. 1953. The Indian as an ecological factor in the northeast forest. *Ecology* 34:329-346
- Denevan, W.M. 1992. The pristine myth: The landscape of the Americas in 1492. *Annals of the Association of American Geographers* 83:369-385.
- Faber-Langendoen, D. editor. 2001. Plant communities of the Midwest: Classification in an ecological context. Association for Biodiversity Information, Arlington, VA. 61 pp. + appendix (705 pp.).



- Glidden, A.C. 1892. Pioneer farming. Michigan Pioneer Historical Collections 18:418-422.
- Harty, F.M. 1986. Exotics and their ecological ramifications. *Natural Areas Journal* 6(4):20-26.
- Herkert, J.R., R.E. Szafoni, V.M. Kleen, and J.E. Schwegman. 1993. Habitat establishment, enhancement and management for forest and grassland birds in Illinois. Division of Natural Heritage, Illinois Department of Conservation, Natural Heritage Technical Publication #1, Springfield, IL, 20 pp.
- Kay, C.E. 1995. Aboriginal overkill and Native burning: Implications for modern ecosystem management. *Western Journal of Applied Forestry* 10:121-126.
- Kost, M.A., D.A. Albert, J.G. Cohen, B.S. Slaughter, R.K. Schillo, C.R. Weber, and K.A. Chapman. 2007. Natural communities of Michigan: Classification and description. Michigan Natural Features Inventory, Report Number 2007-21, Lansing, MI. 314 pp.
- Laubhan, M.K. 1995. Effects of prescribed fire on moist-soil vegetation and macronutrients. *Wetlands* 15:159-66.
- Leach, M.K. and T.J. Givnish. 1996. Ecological determinants of species loss in remnant prairies. *Science* 273:1555-1558.
- NatureServe. 2004. NatureServe Explorer: An online encyclopedia of life [web application]. Version 1.8. NatureServe, Arlington, Virginia. Available <http://www.natureserve.org/explorer>. (Accessed: September 20, 2004).
- Nuzzo, V. 1976. Propagation and planting of prairie forbs and grasses in southern Wisconsin. Pp. 182-189 in *Proceedings of the Fifth Midwest Prairie Conference*. Iowa State University, Ames, Iowa.
- Packard, S. and C.F. Mutel. 1997. The tallgrass restoration handbook for prairies savannas and woodlands. Island Press, Washington D.C. 463 pp.
- Panzer, R.D., D. Stillwaugh, R. Gnaedinger, and G. Derkowitz. 1995. Prevalence of remnant dependence among prairie-and savanna-inhabiting insects of the Chicago region. *Natural Areas Journal* 15:101-116.
- Ratti, J.T. and K.P. Reese. 1988. Preliminary test of the ecological trap hypothesis. *Journal of Wildlife Management* 52:484-491.
- Reich, P.B., M.D. Abrams, D.S. Ellsworth E. L. Kruger and T. J. Tabone. 1990. Fire affects ecophysiology and community dynamics of Central Wisconsin oak forest regeneration. *Ecology* 71:2179-90.
- Sample, D.W. and M.J. Mossman. 1997. Managing habitat for grassland birds: A guide for Wisconsin. Bureau of Integrated Science Services, Department of Natural Resources, Madison, WI. 154 pp.
- Scharrer, E.M. 1972. Relict prairie flora of southwestern Michigan. Pp. 9-12 in J.H. Zimmerman (ed.) *Proceedings of the Second Midwest Prairie Conference*, Madison, WI. 242 pp.
- Schmalzer, P.A. and C.R. Hinkle. 1992. Soil dynamics following fire in *Juncus* and *Spartina* marshes. *Wetlands* 12:8-21.
- Schulenberg, R. 1972. Notes on the propagation of prairie plants. The Morton Arboretum, Lisle IL. 15 pp.
- Smith, L.M. and J.A. Kadlec. 1985. Fire and herbivory in a Great Salt Lake marsh. *Ecology* 66:259-65.
- Steuter, A.A. 1997. Bison. Pp. 339-347 in Packard, S. and C.F. Mutel (eds.), *The tallgrass restoration handbook for prairies savannas and woodlands*. Island Press, Washington D.C. 463 pp.
- Thompson, D.Q. and R.H. Smith. 1970. The forest primeval in the Northeast - a great myth? in *Proceedings of the Tall Timbers Fires Ecology Conference*. 10:255-265.
- Thompson, P.W. 1970. The preservation of prairie stands in Michigan. Pp. 13-14 in J.H. Zimmerman (ed.) *Proceedings of the Second Midwest Prairie Conference*, Madison, WI. 242 pp.
- Thompson, P.W. 1975. The floristic composition of prairie stands in southern Michigan. pp. 317-331 in M.K. Wali (ed.), *Prairie: A multiple view*. The University of North Dakota, Grand Fork, N.D.
- Thompson, P.W. 1983. Composition of prairie stands in southern Michigan and adjoining areas. Pp. 105-111 in R. Brewer (ed.), *Proceedings of the Eighth North American Prairie Conference*.
- Timmins, S.M. 1992. Wetland vegetation recovery after fire: Eweburn Bog, Te Anau, New Zealand. *New Zealand Journal of Botany* 30:383-99.
- Trager, J.C. 1990. Restored prairies colonized by native prairie ants (Missouri, Illinois). *Restoration and Management Notes* 8:104-105.
- Trager, J.C. 1998. An introduction to ants (*Formicidae*) of the tallgrass prairie. *Missouri Prairie Journal* 18:4-8.
- Transeau, E.N. 1935. The prairie peninsula. *Ecology* 16:423-437.
- Viro, P.J. 1974. Effects of forest fire on soil. Pp. 7-45 in T. T. Kozlowski and C. E. Ahlgren (eds.), *Fire and Ecosystems*. Academic Press, New York, NY.
- Vogl, R.J. 1964. The effects of fire on a muskeg in northern Wisconsin. *Journal of Wildlife Management* 28:317-29.



Vogl, R.J. 1974. Effects of fire on grasslands. Pp. 139-94 in T. T. Kozlowski and C. E. Ahlgren (eds.), *Fire and Ecosystems*. Academic Press, New York, NY.

Warners, D.P. 1997. Plant diversity in sedge meadows: Effects of groundwater and fire. Ph.D. dissertation, University of Michigan, Ann Arbor, MI. 231 pp.

Abstract Citation:

Kost, M.A. 2004. Natural community abstract for dry-mesic prairie. Michigan Natural Features Inventory, Lansing, MI. 10 pp.

Updated June 2010.

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Michigan State University is an affirmative-action, equal-opportunity organization.

Funding for this abstract was provided by the Michigan Department of Natural Resources, Landowner Incentive Program, in partnership with the U.S. Fish and Wildlife Service.

